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Genotoxicity and Cytotoxicity of Cadmium Sulfide Nanomaterials to Mice: Comparison Between Nanorods and Nanodots

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Abstract

Cadmium sulfide (CdS) nanomaterials (such as CdS nanodots or nanorods) are widely used in optical, electronic, and biological applications. Large-scale production and use of these materials will likely result in accidental and incidental releases, which raise concerns about their potential environmental and human-health impacts. Most studies on toxicity of Cd-containing nanomaterials have focused on nanodots, and the relative toxicity of Cd-containing nanorods is not well understood. Here, we compared genotoxicity and cytotoxicity of CdS nanorods (30–50 nm diameter, 500–1100 nm length) and cubic CdS nanodots (3–5 nm) in mice by examining total cadmium accumulation in organs, acute toxicity, DNA damage, spermatozoon viability and abnormality, kidney and liver damage, and oxidative stress. Compared with (smaller) nanodots, nanorods resulted in relatively low bioaccumulation, acute toxicity, and damage to spermatozoa and the tested organs. Differences in toxicity between CdS nanodots and nanorods could not be fully explained by differences in their metal ion (Cd²⁺) release patterns, based on control tests with mice gavaged with dissolved CdCl₂ at equivalent concentrations. This underscores that toxicity of metallic nanomaterials could not be solely predicted based either on their elemental composition or on the amount of ions released before receptor intake. Particle morphology (including size) may also need to be considered.

Key words: cadmium sulfide; cytotoxicity; genotoxicity; nanodots; nanorods

Introduction

POTENTIAL HEALTH EFFECTS of engineered nanomaterials have received significant attention (Hardman, 2006; Nel et al., 2006; Fadeel and Garcia-Bennett, 2010). Compared with their bulk counterparts, nanomaterials are often more toxic (Grassian et al., 2006; Pan et al., 2007). For example, a number of toxicological studies using rats have shown that exposure to nanodots induces greater inflammatory and cytotoxic effects than larger-sized particles at equivalent mass concentrations (Grassian et al., 2006; Gillespie et al., 2010; Rossi et al., 2010). Most nanomaterial toxicity studies have focused on nanodots (i.e., particles of various shapes with no predominant dimension, or 0D), and only a few studies have addressed the toxicity

of elongated rod-like (1D) nanoparticles (Magrez *et al.*, 2009; Song *et al.*, 2010) or compared 1D nanomaterials with their 0D counterparts (Chithrani *et al.*, 2006; Pal *et al.*, 2007; Ispas *et al.*, 2009; Simon-Deckers *et al.*, 2009). Therefore, the toxicological implications of nanoparticle morphology (including shape and size) are not fully understood.

Cadmium-containing nanomaterials are widely used in optical, electronic, and biological applications (Nirmal et al., 1996; Agarwal et al., 2005; Bowers et al., 2005). Accordingly, concern about their potential environmental and human-health impacts has increased (Rzigalinski and Strobl, 2009). It has been argued that toxic effects of CdSe are largely linked to Cd²⁺ released from the CdSe surface (Kirchner et al., 2005; Mahendra et al., 2008; Su et al., 2010). Cd² released from quantum dots is bactericidal, and it may also induce production of intracellular reactive oxygen species (ROS), resulting in oxidative stress (Dailianis et al., 2005; Yang et al., 2007a, 2007b). Haque et al. (2013) concluded that oxidative stress caused by higher production of ROS was an important factor in the pathogenicity of CdSe/CdS-MPA nanodots to mice. Hossain and Mukherjee (2013) reported that ROS levels in both Escherichia coli and HeLa cells increased on exposure to CdS nanodots; in addition, HeLa cells

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exhibited altered morphology with condensed and fragmented nuclei. Such overloads of ROS in mammalian cells can damage membrane lipids, proteins, and DNA (Finkel and Holbrook, 2000). Li *et al.* (2009) argued that both increased intracellular ROS and Cd²⁺ release are possible mechanisms for the cytotoxicity of CdS quantum dots. They also suggested that with the increase of quantum dot concentration, the toxicity mechanism changes from intracellular oxidative stress to Cd²⁺ release.

To date, the relative toxicity of cadmium-containing nanorods is not well understood. In particular, it is unknown whether differences in acute toxicity and genotoxicity between Cd-containing nanodots and nanorods are mainly due to differences in their Cd²⁺-release characteristics. In this study, we used mice to compare genotoxicity and cytotoxicity of CdS 1D nanorods versus CdS 0D nanodots, using CdCl₂ salt as a control for the effect of released Cd²⁺. Acute toxicity of the two CdS nanoproducts was compared along with bioaccumulation of total Cd in different organs, including kidney, liver, and spermary. A comet assay with mice lymphocytes was also conducted to evaluate DNA damage by means of single-cell gel electrophoresis. Damage to spermatozoa, liver, and kidney after exposure to nanodots or nanorods was also compared.

Materials and Methods

Reagents and animals

All chemical reagents were purchased from Beijing Chemical Reagent Ltd., China, and used without further purification. Superoxide dismutase (SOD) and malondialdehyde (MDA) kits were purchased from Beyotime Institute of Biotechnology, China. Kunming mice (17–22 g/mouse) were obtained from the Center for Experimental Animals of North China Coal Medical University. The mice were maintained at 22°C for one week before use in the experiments.

Synthesis and characterization of CdS nanomaterials

CdS nanodots were synthesized by adding 200 mL of 0.2 M Na₂S to 200 mL of 0.2 M CdCl₂ solution under agitation (85-2, Gongyi Kehua) (700 rpm). To synthesize CdS nanorods, CdCl₂ (0.3 g) and thioacetamide (0.4 g) were added to a 30-mL Teflon-lined stainless steel autoclave, followed by the addition of 2 mL of distilled water and 18 mL of ethylenediamine. The autoclave was placed in an oven at 150°C for 24 h, and then allowed to cool to room temperature. The yellowish product was collected and washed with distilled water and ethanol.

Sizes and shapes of all the samples were examined using a Hitachi 3500 scanning electron microscope (SEM), as well as a JEOL-2010 high-resolution transmission electron microscope (TEM) that was operated at an acceleration voltage of 200 kV. The crystal structure was characterized with a Rigaku D/Max-2500 X-ray diffractometer (XRD) employing Cu K α radiation, λ =1.54056 Å. Hydrodynamic diameters were measured by dynamic light scattering, and ζ potentials were measured by electrophoretic mobility, on a Malvern Zetasizer (ZETAPALS/BI-200SM, Brookhaven).

Characterization of CdS dissolution at different pH values

To characterize dissolution kinetics, CdS nanorods or nanodots were suspended in saline (0.9% NaCl) at 10 g/L. The

pH was adjusted with HCl or NaOH. After agitation for designated time intervals, 10 mL of the suspensions was taken out and centrifuged at 25,750 g for 10 min to separate the CdS nanomaterial and the solution (Chen *et al.*, 2012). The supernatant was withdrawn, and the Cd²⁺ concentration in the solution was measured with inductively coupled plasma (ICP; ICP-9000, Jarrell-Ash). For each CdS product, the dissolution kinetics was examined at pH 2.0 and 7.0, respectively.

Acute toxicity

Acute toxicity of CdS nanodots, CdS nanorods, and CdCl₂ salt was tested. The study population of 120 mice were randomly divided into 12 groups (each containing 10 mice). After an absolute diet for 12 h, CdS nanoproducts or CdCl₂ salt were gavaged to the mice. Gavagation was performed daily for 2 weeks. The daily doses used were: CdS nanodots, 215, 464, 1000, and 2150 mg/kg; CdS nanorods, 1000, 2150, 4640, and 10,000 mg/kg; and CdCl₂, 46.4, 100, 215, and 464 mg/kg. The median lethal dose (LD₅₀) was calculated using Horn's equation (Horn *et al.*, 1956).

Accumulation of cadmium in mice

To test cadmium accumulation, 112 mice were randomly divided into seven groups (each containing 16 mice). Groups 1 and 2 were gavaged with high doses of CdS nanodots or nanorods; the concentrations of CdS nanomaterials were 10 mg/mL in saline (equivalent to a dose of 200 mg/kg per day). Groups 3 and 4 were gavaged with low doses of CdS nanodots or nanorods; the concentration of CdS nanomaterials were 5 mg/mL in saline (100 mg/kg per day). Groups 5 and 6 were gavaged with 12.7 g/L (high dose, 254 mg/kg per day) or 6.35 g/L (low dose, 127 mg/kg per day) of CdCl₂ saline solution (the total mass of Cd in the solutions was equal to the respective mass of Cd involved in the experiments using CdS nanomaterials). Group 7 was used as the control and was gavaged with a saline containing 0.1 mg/L CdCl₂ (equivalent to a dose of $2 \mu g/kg$ per day); this concentration of Cd²⁺ is slightly higher than the highest concentrations of dissolved Cd²⁺ from the CdS nanorods and nanodots as measured in the CdS dissolution experiments. After 18 or 35 days, eight mice from each group were dissected, and target organs, including liver, kidney, and spermary, were obtained. The target organs were liquefied with 5 mL nitric acid and 1 mL hydrogen peroxide, microwave digested (Multiwave 3000, Anton Paar GmbH), and diluted to a final volume of 10 mL. Concentrations of total Cd in tissues were measured with an atomic absorption spectrum (SP-3520AA).

Comet assay for DNA damage

Forty mice were randomly divided into four groups (each containing 10 mice). Each group was gavaged with saline containing 10 g/L CdS nanodots (equivalent to a dose of 200 mg/kg per day), 10 g/L CdS nanorods (200 mg/kg per day), 12.7 g/L of CdCl₂ (254 mg/kg per day), or 0.1 mg/L CdCl₂ (control, 2 μ g/kg per day), respectively. The comet assay experiments were carried out after 35 days following the standard protocol (Olive and Banath, 2006). Cells were analyzed at 400× magnification with a Komet 3.1 Image Analysis System (Kinetic Imaging) using a Leica DMLB fluorescence microscope.

Sperm quantity and quality

Thirty male mice were randomly divided into three groups (each containing 10 mice), and they were gavaged with 10 g/L CdS nanodots, 10 g/L CdS nanorods, and saline with 0.1 mg/L CdCl₂ (control), respectively. After 35 days, the epididymis were removed and cut into pieces in physiological saline, and then filtrated. The filtrated saline solution was stained with 1% eosin solution for 15 min, washed with distilled water, and left to dry in the air (Bai *et al.*, 2010; Elmazoudy *et al.*, 2011). The samples were analyzed with a microscope, and the spermatozoon sperm quantity and quality was characterized from 5000 integral spermatozoa.

Damage to kidney and liver, and SOD and MDA assays

Forty mice were randomly divided into four groups (each containing 10 mice). Each group was gavaged daily with $10\,\text{g/L}$ CdS nanodots (equivalent to $200\,\text{mg/kg}$), $10\,\text{g/L}$ CdS nanorods ($200\,\text{mg/kg}$), $12.7\,\text{g/L}$ CdCl₂ ($254\,\text{mg/kg}$), and $0.1\,\text{mg/L}$ CdCl₂ (control, $2\,\mu\text{g/kg}$), respectively. The mice were killed by vertebral dislocation after 16 days, and their kidneys and livers were fixed by formalin. The fixed samples were sliced and stained with hematoxylin and eosin, and then histopathology examinations were performed. The SOD and MDA levels in mice blood serum were assayed using SOD and MDA assay kits S0081 and S0131 (Beyotime Institute of Biotechnology), respectively.

Results and Discussion

Characteristics of CdS nanodots and nanorods

TEM and SEM images of the as-prepared CdS products are shown in Fig. 1. The dimensions of CdS nanodots ranged from

3 to 5 nm (Fig. 1a), whereas CdS nanorods had diameters ranging from 30 to 50 nm and lengths from 500 to 1100 nm (Fig. 1b). The XRD patterns (Fig. 1c,d) show that the nanodots had a cubic structure (a = 5.818 Å, JCPDS: 10-0454) and the nanorods had a hexagonal structure (a = 4.141 Å, c = 6.72 Å, JCPDS: 41-1049). The ζ potential values (in ethanol) were 4.0 mV for CdS nanodots and 18.1 mV for nanorods.

Significant aggregation of CdS nanodots occurred in the saline exposure medium; the particle-sized distribution exhibited a large peak at 4914 nm and a small peak at 850 nm (Fig. 2a). Both differences in nanoparticle size and extent of aggregation represent potential confounding effects for the effective dose and distribution of nanomaterials delivered, which could affect the physiological response in mice. One additional confounding factor is that the nanomaterial aggregates that form in well-defined exposure media are not representative of the more complex and more difficult to characterize hetero-aggregates which form *in vivo* (Albanese and Chan, 2011). Despite such common confounding factors, it is important to explore how initial morphological differences in CdS nanoparticles affect their toxicity to different mammalian systems.

Dissolution of CdS nanodots and nanorods under different pH values

Both CdS nanodots and CdS nanorods released more Cd^{2+} at pH 2.0 than at pH 7.0, and dissolution leveled off after \sim 60 h (Fig. 2b, c). The smaller nanodots (with thus higher specific surface area) released more Cd^{2+} than the nanorods. The concentration of dissolved Cd^{2+} reached 30 g/L for CdS nanodots, compared with 1.5 μ g/L for CdS nanorods (for pH 7.0). For both nanoparticle suspensions, the released Cd^{2+} accounted for a relatively small fraction of the added total Cd. For example, at pH 2.0, the mass of Cd released was 0.0005%

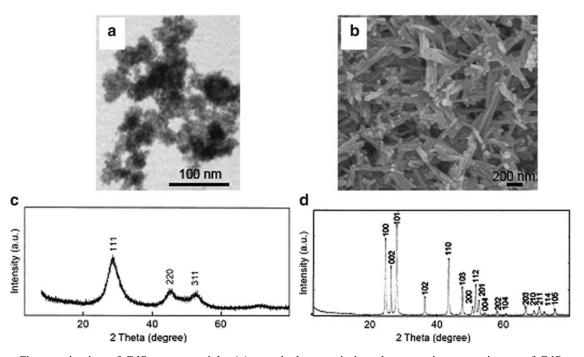


FIG. 1. Characterization of CdS nanomaterials: (a) a typical transmission electron microscope image of CdS nanodots; (b) a typical scanning electron microscope image of CdS nanorods; (c) X-ray diffractometer (XRD) pattern of CdS nanodots; and (d) XRD pattern of CdS nanorods.

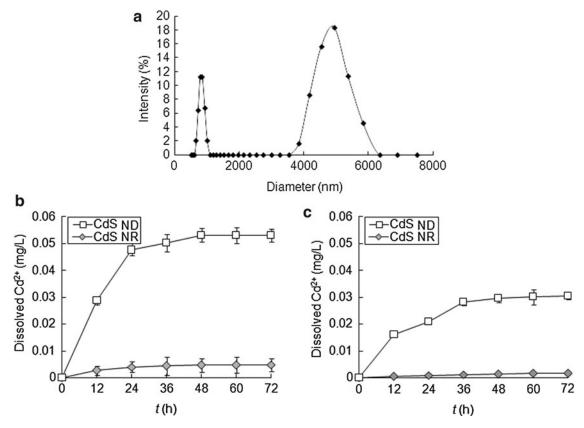


FIG. 2. (a) Intensity-weighted particle-sized distribution of CdS nanodots. Differences in dissolution kinetics between CdS nanodots and nanorods are also depicted, at (b) pH 2.0 and (c) pH 7.0. ND, nanodots; NR, nanorods.

of the total amount of Cd contained in CdS nanodots, and 0.00003% for CdS nanorods.

Acute toxicity of CdS nanodots and nanorods

The LD_{50} values of CdS nanodots and CdS nanorods were 767 and 7203 mg/kg per day (Fig. 3), respectively; whereas the LD_{50} value of CdCl₂ was lower at 137 mg/kg per day. The value corresponding to CdS nanodots is lower than that reported for bulk CdS material for mice (1166 mg/kg per day, according to the material safety data sheet), suggesting higher toxicity of these nanoparticles. On the

other hand, the LD_{50} for CdS nanorods is significantly higher than that of the bulk CdS. It should be noted that the LD_{50} value of CdS nanodots was only 5.6 times higher than the LD_{50} value of CdCl₂, even though CdS nanodots showed a limited dissolution potential (0.0005% of total Cd present) in the exposure medium (Fig. 3); whereas CdCl₂ was completely dissolved. This suggests that the toxicity of Cd-containing nanomaterials cannot be simply predicted based on the amount of Cd²⁺ released before intake. Obviously, the bioavailable Cd that contributes to toxicity encompasses a greater fraction than Cd²⁺ ions dissolved in the exposure medium.

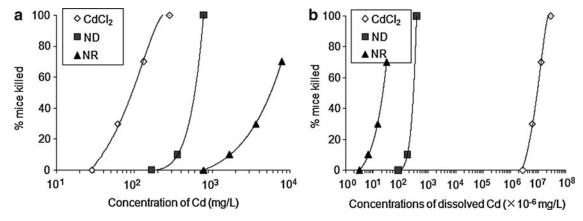


FIG. 3. Percentage of mice killed versus (a) total concentration of Cd and (b) concentration of dissolved Cd (calculated from the dissolution curves).

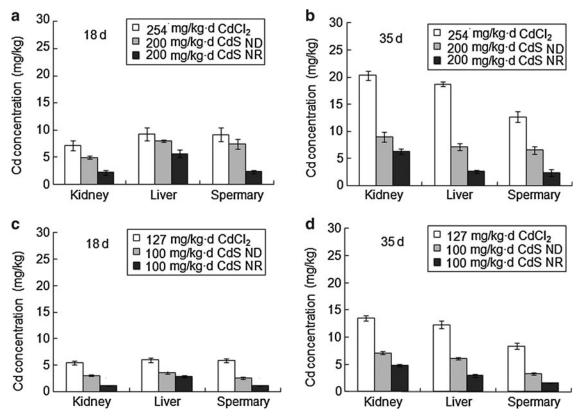


FIG. 4. Accumulation of Cd in target organs: mice treated for 18 (a) and 35 days (b) with a high dose (200 mg/kg) of CdS nanomaterials; mice treated for 18 (c) and 35 days (d) with a low dose (100 mg/kg) of CdS nanomaterials.

Accumulation of cadmium in target organs

Cd concentrations in all test organs (kidney, liver, and spermary) were generally considerably higher for the mice treated with CdS nanodots than for those treated with equivalent doses of CdS nanorods, but not as high for mice treated with equivalent concentrations of CdCl₂ solution (Fig. 4). No accumulation was observed for the control (mice treated with 0.1 mg/L CdCl₂). Accumulation in all organs of the mice treated with a low dose (100 mg/kg per day) was about one-half of that for the mice receiving a high dose (200 mg/kg per day: Fig. 4a.c), indicating a relatively linear relationship between dose and accumulation. The relative concentrations of Cd in the three target organs changed over time. For example, accumulation in the liver decreased from day 18 to day 35 (Fig. 4b,d), suggesting excretion. For a given nanoparticle type, the accumulation rate differed among different organs as previously reported (Hauck et al., 2010).

Even though the highest total Cd accumulation was observed in mice treated with CdCl₂ solution, the differences in

accumulation between different treatments were relatively small (accumulation in mice treated with CdCl₂ was 1.1 to 2.5 times greater than in mice gavaged with CdS nanodots, and 1.6 to 7.3 times greater than in mice gavaged with CdS nanorods), considering that (unlike the CdCl₂ treatment) only a small fraction of CdS nanodots and nanorods was present as dissolved Cd²⁺ during exposure. Moreover, differences in Cd accumulation between mice gavaged with CdS nanodots and nanorods were rather small compared with the large differences in their Cd²⁺ release patterns. Thus, the differences in total Cd bioaccumulation between the two types of nanopaticles cannot be explained solely based on differences in their capability to release Cd²⁺ ions.

Genotoxicity of CdS nanodots and nanorods

The genotoxic effects of CdS nanodots and nanorods were examined with single-cell gel electrophoresis (comet assay; Table 1). Damage to DNA, as indicated by the significance of

TABLE 1. RESULTS OF SINGLE-CELL GEL ELECTROPHORESIS

Samples	Number of cells observed	Tailed DNA (%)	Tail length (µm)
Negative control	500	0.28±0.13	1.2±0.1
Nanorods	500	37±6	6.2±0.2
Nanodots	500	48±8	19±0
Positive control	500	55±10	18±0

Table 2. Results of Spermatozoon Abnormality and Normality

Samples		Number of spermatozoa observed	Abnormality (%)	Normality (%)
Negative control	10	5000	0.15 ± 0.11	97±3
Nanorods Nanodots	10 10	5000 5000	0.89 ± 0.52 2.3 ± 0.9	77 ± 8 1.8 ± 3.3

tailing of lymphocytes, showed that genotoxicity was greater for mice treated with CdCl₂ solution, followed by mice treated with CdS nanodots and then mice treated with CdS nanorods. The percentage of tailing for the normal lymphocytes was only $0.28\% \pm 0.13\%$ (Table 1), compared with $37\% \pm 6\%$ (with a tail length of $6.2 \pm 0.2~\mu m$) for the mice treated with CdS nanorods, $48\% \pm 8\%$ (with tail length of $19.0 \pm 0.0~\mu m$) for the mice treated with CdS nanodots, and $55\% \pm 10\%$ for the mice treated with CdCl₂ solution. These data corroborate the small differences in effects observed between the mice treated with CdS nanodots and those treated with nanorods, reinforcing the inference that the dissolution potential of CdS

nanomaterials in the exposure medium was not the most critical factor for controlling toxicity.

To further evaluate the differences in genotoxicity between CdS nanodots and nanorods, we examined the spermary damnification of the mice (Table 2; Supplementary Fig. S1). Compared with the control group, an obvious decrease in the number of spermatozoa was observed for mice gavaged with CdS nanorods or CdS nanodots. The sperm quantity and quality of the spermatozoa also varied significantly among different samples (Table 2; Supplementary Fig. S1.). Spermatozoon abnormality for the control group was $0.15\% \pm 0.11\%$, compared with $0.89\% \pm 0.52\%$ for the mice treated with CdS

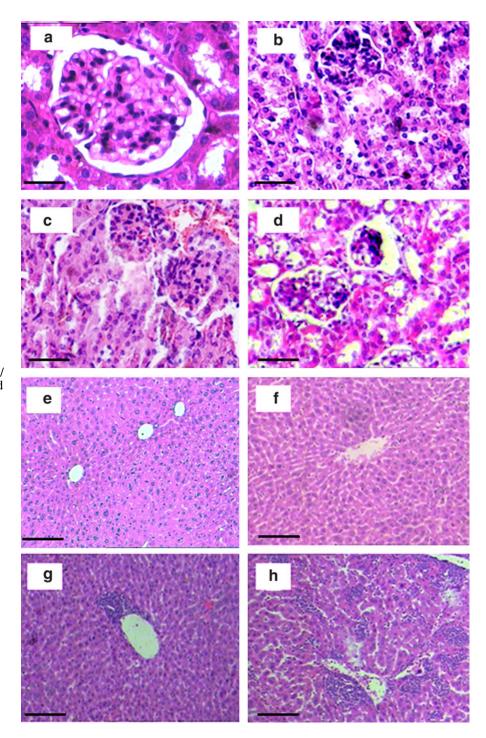


FIG. 5. Histological sections of mice renal (**a–d**) and liver tissues (**e–h**), on treatment for 16 days with CdS nanomaterials: (**a, e**) control; (**b, f**) treated with 200 mg/kg per day nanorods; (**c, g**) treated with 200 mg/kg per day nanodots; (**d, h**) treated with 254 mg/kg per day of CdCl₂. Scale bar = 50 μm.

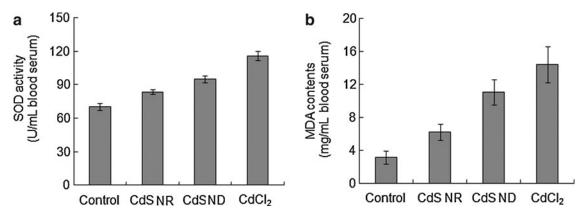


FIG. 6. Oxidative stress exerted by CdS nanomaterials in blood serum. Effects of different treatments for 16 days on: (a) superoxide dismutase (SOD) activities (n=10 per group); (b) malondialdehyde (MDA) contents (n=10 per group).

nanorods and with $2.3\% \pm 0.9\%$ for the mice treated with CdS nanodots. More profound effects were observed for normal spermatozoa, which decreased from $97.0\% \pm 3.0\%$ for the control to $77.0\% \pm 8.0\%$ for CdS nanorods and to $1.8\% \pm 3.3\%$ for the mice treated with CdS nanodots.

Kidney and liver damage

Compared with the normal renal histopathology image of the control group (Fig. 5a), mice gavaged with CdS nanorods (Fig. 5b) retained glomerulars integrity, but renal interstitial engorgement and edema appeared and leucocytes became more abundant. The histopathology image of the mice gavaged with CdS nanodots (Fig. 5c) showed slight glomerular damage, which is conducive to inflammation and injury of the renal tissues. For the mice gavaged with CdCl₂ (Fig. 5d), the glomerulars were significantly damaged and many lymphocygtes were observed. Cd-induced damage was irreversible even after exposure ended. These results corroborate that CdS nanorods were less toxic than CdS nanodots.

A similar trend was observed for effects on the liver. Normal liver cells (Fig. 5e) were radioactively aligned around lobule central vein and had polygon shapes, and liver sinusoidal exited among hepatic cords. For the mice gavaged with CdS nanorods (Fig. 5f), a few inflammatory cells appeared around the lobule central vein; the hepatic cords showed radioactive and regular arrangement; and only a small quantity of liver cells were edemas. In contrast, many inflammatory cells were observed around the lobule central vein for mice treated with CdCl₂ (Fig. 5h) or CdS nanodots (Fig. 5g). These treatments also resulted in hepatic and hepatocelluar vacuolation or steatosis in liver cells. Thus, CdS nanorods caused less liver damage than CdS nanodots.

Oxidative stress

SOD and MDA data were obtained to evaluate oxidative stress (Fig. 6). Similar to other toxicity indicators discussed earlier, both SOD and MDA levels indicated a greater effect for the mice gavaged with CdCl₂ solution, followed by the mice gavaged with CdS nanodots and then the mice gavaged with CdS nanorods. The increased SOD and MDA levels in the blood serum were likely caused by excess ROS production, and implicated oxidative stress as a cause of the observed damage to kidney and liver tissues and spermatozoon

abnormality as observed in other studies (Bai *et al.*, 2010; Haque *et al.*, 2013). This suggests that valuable fundamental insights could be provided by immunotoxicology studies which consider organismal response as a function of particle morphology, including size, shape, and extent of aggregation.

Conclusions

CdS nanodots and CdS nanorods exhibited markedly different Cd bioaccumulation, genotoxicity, and cytotoxicity in mice that cannot be fully explained by the differences in their Cd²⁺⁻release patterns. The mechanisms and factors controlling bioaccumulation and toxicity of metallic nanomaterials are rather complex [e.g., released ions can be bound by common ligands that decrease their availability while remaining in solution (Xiu *et al.*, 2012)], and their toxicity cannot be predicted based solely on their elemental composition. Particle morphology (including size, shape, and extent of aggregation) should also be considered. This study also underscores the importance of averting unintended releases to the environment and the need for further research on how environmental and *in vivo* transformations (and associated matrix effects) affect metal speciation and the resulting biological interactions.

Acknowledgments

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Author Disclosure Statement

No competing financial interests exist.

References

Agarwal, R., Barrelet, C.J., and Lieber, C.M. (2005). Lasing in single cadmium sulfide nanowire optical cavities. *Nano Lett.* 5, 917.

Albanese, A., and Chan, W.C.W. (2011). Effect of gold nanoparticle aggregation on cell uptake and toxicity. *ACS Nano* 5, 5478.

Bai, Y.H., Zhang, Y., Zhang, J.P., Mu, Q.X., Zhang, W.D., Butch, E.R., Snyder, S.E., and Yan, B. (2010). Repeated

administrations of carbon nanotubes in male mice cause reversible testis damage without affecting fertility. *Nat. Nanotechnol.* 5, 683.

- Bowers, M.J., McBride, J.R., and Rosenthal, S.J. (2005). White-light emission from magic-sized cadmium selenide nanocrystals. *J. Am. Chem. Soc.* 127, 15378.
- Chen, P., Powell, B.A., Mortimer, M., and Ke, P.C. (2012). Adaptive interactions between zinc oxide nanoparticles and *Chlorella* sp. *Environ. Sci. Technol.* 46, 12178.
- Chithrani, B.D., Ghazani, A.A., and Chan, W.C. (2006). Determining the size and shape dependence of gold nanoparticle uptake into mammalian cells. *Nano Lett.* 6, 662.
- Dailianis, S., Piperakis, S.M., and Kaloyianni, M. (2005). Cadmium effects on ROS production and DNA damage via adrenergic receptors stimulation: Role of Na⁺/H⁺ exchanger and PKC. Free Radic. Res. 39, 1059.
- Elmazoudy, R.H., Attia, A.A., and El-Shenawy, N.S. (2011). Protective role of propolis against reproductive toxicity of chlorpyrifos in male rats. *Pestic. Biochem. Physiol.* 101, 175.
- Fadeel, B., and Garcia-Bennett, A.E. (2010). Better safe than sorry: Understanding the toxicological properties of inorganic nanoparticles manufactured for biomedical applications. *Adv. Drug Deliv. Rev.* 62, 362.
- Finkel, T., and Holbrook, N.J. (2000). Oxidants, oxidative stress and the biology of ageing. *Nature* 408, 239.
- Gillespie, P.A., Kang, G.S., Elder, A., Gelein, R., Chen, L., Moreira, A.L., Koberstein, J., Tchou-Wong, K.M., Gordon, T., and Chen, L.C. (2010). Pulmonary response after exposure to inhaled nickel hydroxide nanoparticles: Short and long-term studies in mice. *Nanotoxicology* 4, 106.
- Grassian, V.H., O'Shaughnessy, P.T., Adamcakova-Dodd, A., Pettibone, J.M., and Thorne, P.S. (2006). Inhalation exposure study of titanium dioxide nanoparticles with a primary particle size of 2 to 5 nm. *Environ. Health Perspect.* 115, 397.
- Haque, M.M., Im, H.Y., Seo, J.E., Hasan, M., Woo, K., and Kwon, O.S. (2013). Acute toxicity and tissue distribution of CdSe/CdS-MPA quantum dots after repeated intraperitoneal injection to mice. J. Appl. Toxicol. 33, 940.
- Hardman, R. (2006). A toxicologic review of quantum dots: toxicity depends on physicochemical and environmental factors. *Environ. Health Perspect.* 114, 165.
- Hauck, T.S., Anderson, R.E., Fischer, H.C., Newbigging, S., and Chan, W.C. (2010). *In vivo* quantum-dot toxicity assessment. *Small* 6, 138.
- Horn, H.J., Laboratories, H., and Virginia, F.C. (1956). Simplified LD₅₀ (or ED₅₀) calculations. *Biometrics* 12, 311.
- Hossain, S.T., and Mukherjee, S.K. (2013). Toxicity of cadmium sulfide (CdS) nanoparticles against *Escherichia coli* and HeLa cells. *J. Hazard. Mater.* 260,1073.
- Ispas, C., Andreescu, D., Patel, A., Goia, D.V., Andreescu, S., and Wallace, K.N. (2009). Toxicity and developmental defects of different sizes and shape nickel nanoparticles in zebrafish. *Environ. Sci. Technol.* 43, 6349.
- Kirchner, C., Liedl, T., Kudera, S., Pellegrino, T., Munoz-Javier, A., Gaub, H.E., Stolzle, S., Fertig, N., and Parak, W.J. (2005). Cytotoxicity of colloidal CdSe and CdSe/ZnS nanoparticles. *Nano Lett.* 5, 331.
- Li, K.G., Chen, J.T., Bai, S.S., Wen, X., Song, S.Y., Yu, Q., Li, J., and Wang, Y.Q. (2009). Intracellular oxidative stress and

- cadmium ions release induce cytotoxicity of unmodified cadmium sulfide quantum dots. *Toxicol. In Vitro* 23, 1007.
- Magrez, A., Horvath, L., Smajda, R., Salicio, V., Pasquier, N., Forro, L., and Schwaller, B. (2009). Cellular toxicity of TiO₂-based nanofilaments. *ACS Nano* 3, 2274.
- Mahendra, S., Zhu, H., Colvin, V.L., and Alvarez, P.J. (2008).Quantum Dot weathering results in microbial toxicity. *Environ. Sci. Technol.* 42, 9424.
- Nel, A., Xia, T., Madler, L., and Li, N. (2006). Toxic potential of materials at the nanolevel. *Science* 311, 622.
- Nirmal, M., Dabbousi, B.O., Bawendi, M.G., Macklin, J.J., Trautman, J.K., Harris, T.D., and Brus, L.E. (1996). Fluorescence intermittency in single cadmium selenide nanocrystals. *Nature* 383, 802.
- Olive, P.L., and Banath, J.P. (2006). The comet assay: a method to measure DNA damage in individual cells. *Nat. Protoc.* 1, 23.
- Pal, S., Tak, Y.K., and Song, J.M. (2007). Does the antibacterial activity of silver nanoparticles depend on the shape of the nanoparticle? A study of the gram-negative bacterium *Escherichia coli. Appl. Environ. Microbiol.* 73, 1712.
- Pan, Y., Neuss, S., Leifert, A., Fischler, M., Wen, F., Simon, U., Schmid, G., Brandau, W., and Jahnen-Dechent, W. (2007). Size-dependent cytotoxicity of gold nanoparticles. *Small* 3, 1941.
- Rossi, E.M., Pylkkanen, L., Koivisto, A.J., Vippola, M., Jensen, K.A., Miettinen, M., Sirola, K., Nykasenoja, H., Karisola, P., Stjernvall, T., Vanhala, E., Kiilunen, M., Pasanen, P., Makinen, M., Hameri, K., Joutsensaari, J., Tuomi, T., Jokiniemi, J., Wolff, H., Savolainen, K., Matikainen, S., and Alenius, H. (2010). Airway exposure to silica-coated TiO₂ nanoparticles induces pulmonary neutrophilia in mice. *Toxicol Sci* 113, 422.
- Rzigalinski, B.A., and Strobl, J.S. (2009). Cadmium-containing nanoparticles: Perspectives on pharmacology and toxicology of quantum dots. *Toxicol. Appl. Pharmacol.* 238, 280.
- Simon-Deckers, A., Loo, S., Mayne-L'ermite, M., Herlin-Boime, N., Menguy, N., Reynaud, C., Gouget, B., and Carriere, M. (2009). Size-, Composition- and shape-dependent toxicological impact of metal oxide nanoparticles and carbon nanotubes toward bacteria. *Environ. Sci. Technol.* 43, 8423.
- Song, M.M., Song, W.J., Bi, H, Wang, J., Wu, W.L., Sun, J, and Yu, M. (2010). Cytotoxicity and cellular uptake of iron nanowires. *Biomaterials* 31, 1509.
- Su, Y., Hu, M., Fan, C., He, Y., Li, Q., Li, W., Wang, L.H., Shen, P., and Huang, Q. (2010). The cytotoxicity of CdTe quantum dots and the relative contributions from released cadmium ions and nanoparticle properties. *Biomaterials* 31, 4829.
- Xiu, Z.M., Zhang, Q.B., Puppala, H.L., Colvin, V.L., and Alvarez, P.J. (2012). Negligible particle-specific antibacterial activity of silver nanoparticles. *Nano Lett.* 12, 4271.
- Yang, P.M., Chen, H.C., Tsai, J.S., and Lin, L.Y. (2007a). Cadmium induces Ca²⁺-dependent necrotic cell death through calpain-triggered mitochondrial depolarization and reactive oxygen species-mediated inhibition of nuclear factor-KB activity. *Chem. Res. Toxicol.* 20, 406.
- Yang, R.S., Chang, L.W., Wu, J.P., Tsai, M.H., Wang, H.J., Kuo, Y.C., Yeh, T.K., Yang, C.S., and Lin, P. (2007b). Persistent tissue kinetics and redistribution of nanoparticles, quantum dot 705, in mice: ICP-MS quantitative assessment. *Environ. Health Perspect.* 115, 1339.